Version: December 6, 2022

# Recommendations for a Conowingo Dredging BMP Expert Panel Establishment Group

Prepared for the Chesapeake Bay Program Partnership's Water Quality Goal Implementation Team & Modeling Workgroup

#### **Background**

Since July 2021, the Conowingo WIP (CWIP) Steering Committee has worked with the Water Quality Goal Implementation Team (WQGIT) and Modeling Workgroup to consider potential modeling approaches that could serve to estimate nutrient and sediment reductions that may be associated with dredging collected sediment from behind the Conowingo dam. See Attachment A. Maryland has been the lead for these partnership discussions and has engaged with both groups over the course of 2022. Through these discussions, the WQGIT and Modeling Workgroup have determined that a focused group of individuals with relevant expertise is needed to investigate the modeling approach and modeling options. The EPEG's conclusions would determine the need for a BMP expert panel that would, if formed, fulfill the detailed requirements of the BMP Review Protocol in order to establish a new BMP for dredging that could be simulated in CAST scenarios.

#### Charge

The Conowingo Dredging Expert Panel Establishment Group (EPEG) is proposed to:

- Review and evaluate the proposed Conowingo expert panel approach (attached) provided by the CWIP Steering Committee to the WQGIT and Modeling WG to include recommendations on the modeling approach.
- Make recommendations as to whether the proposed modeling or alternative approach would provide enough evidence for an Expert Panel to evaluate the nutrient reduction effectiveness of Conowingo dredging.
  - If an expert panel process is recommended based on the proposed modeling or an alternative approach, then:
    - Make recommendations on any needed adjustments to the proposed approach to inform a more comprehensive scope for the EP effort. This includes:
      - Identify priority tasks for an Expert Panel (EP),
      - Recommend areas of expertise that should be included on the EP, and
      - Draft the EP's charge for the review process
    - Identify the need for funding commitments for the expert panel process.
  - o If an EP is not recommended, provide justification for not convening an EP

To fulfill the above assignment, it is intended for the EPEG to consist of prominent regional experts with existing knowledge sufficient to inform, deliberate and reach conclusions without new or extensive literature reviews, or new intensive analysis. The EPEG is encouraged to build

on existing information in their evaluation of the merits of the proposed modeling approach and their determination to recommend - or not - the formation of an expert panel.

The composition of the proposed EPEG will include no more than 5-7 individuals with, collectively, the following expertise.

- Estuarine and reservoir sediment transport
- Estuarine and reservoir biogeochemistry
- Reservoir hydrodynamics
- Sediment biogeochemistry
- Dredging processes and effects
- Fate, transport, and handling of dredged materials
- Familiarity with scientific literature and previous or ongoing studies of the Lower Susquehanna reservoirs, such as the Lower Susquehanna River Watershed Assessment

Additional participants may include representatives from the CWIP Steering Committee, as well as CBP or jurisdiction agency staff with relevant expertise, such as:

- BMP tracking/reporting
- Chesapeake TMDL modeling
- Previous CBP work on the lower Susquehanna reservoirs
- CWIP processes

Support to the EPEG will also include administrative support to facilitate and schedule meetings, create agendas and take minutes.

If an expert panel is recommended, the EPEG members are not committed to serve on that new panel; nor are they excluded from it. If one or more EPEG members meets the recommended expertise for a new panel, EPEG members may consider volunteering for continued service on the full panel, but a decision on panel membership will come after the EPEG has completed its work and disbanded, with the timing dependent on the support available to a full panel.

Table 1. Proposed Expert Panel Establishment Group membership and affiliations.

| NOTE: Prospective members for sake of discussion, not yet contacted or committed |   |  |  |  |  |
|--|---|--|--|--|--|
| (Prospective) Member   | Affiliation   |  |  |  |  |
| Larry Sanford  | University of Maryland Center for Environmental Science |  |  |  |  |
| Jeff Cornwell  | University of Maryland Center for Environmental Science |  |  |  |  |
| Carl Cerco   | (Retired) US Army Corps of Engineers                    |  |  |  |  |
| Joel Blomquist   | USGS  |  |  |  |  |
| Andy Miller  | University of Maryland – Baltimore County               |  |  |  |  |
| Kathy Boomer   | Foundation for Food and Agriculture Research            |  |  |  |  |
| Jeremy Testa   | University of Maryland Center for Environmental Science |  |  |  |  |
|  |   |  |  |  |  |
| Other EPEG Participants and Support Staff  |   |  |  |  |  |
| Matt Rowe MDE, CWIP Steering Committee Co-Chair                                  |   |  |  |  |  |

| Andy Thuman   | HDR                                      |
|---------------|--|
| Lew Linker    | EPA CBPO; Modeling Workgroup Coordinator |
| Gary Shenk    | USGS, CBP                                |
| Jeremy Hanson | CRC, CBP; WQGIT Coordinator              |

#### Method, Timeline and Deliverables

The EPEG is anticipated to convene virtually or in hybrid form as soon as possible in the new year (2023) to deliberate on and complete its assigned charge within 3-6 months. They will be asked to consider readily available resources and studies such as those referenced in the attached CWIP Steering Committee memo and the <u>current pilot study</u> being conducted by Maryland. Limited review of literature may be needed to supplement the EPEG's deliberations, but comprehensive or systematic literature reviews are not expected for this group.

The EPEG will deliver a memo that describes their consensus conclusions in response to requested tasks above. The WQGIT will consider available next steps based on the EPEG's conclusions, and if an EP is recommended then the partnership will consider support options based on the EPEG's suggested funding needs to support a panel.

## Resources and Information for Consideration by the EPEG

Lower Susquehanna River Watershed Assessment: https://dnr.maryland.gov/waters/bay/pages/lsrwa/final-report.aspx

Northgate Environmental Management, Inc. 2021. Sediment Characterization Study Report (RE01) Conowingo Sediment Characterization and Innovative Reuse and Beneficial Use Pilot Project. Project ID No. 1-18-3-21-8R. Cecil and Harford Counties, Maryland. <a href="https://mde.maryland.gov/programs/marylander/Documents/Final%20Sediment%20Characterization%20Report.pdf">https://mde.maryland.gov/programs/marylander/Documents/Final%20Sediment%20Characterization%20Report.pdf</a>

#### Related modeling documentation--

- Model Flow Chart summarizing model inputs and outputs for the following:
  - Sediment transport model: <u>Estuarine and Coastal Ocean Model (ECOM)</u>
     Sediment Manual v1.3
  - o Sediment flux model: <u>RCA (Row-Column AESOP) User's Guide for release</u> 3.0

# ATTACHMENT A: CWIP Steering Committee memo



# **Conowingo WIP Steering Committee**

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## May 3, 2022

Ed Dunne (WQGIT Chair), District of Columbia Department of Energy & Environment (DOEE)
Suzanne Trevena (WQGIT Vice Chair), U.S. Environmental Protection Agency - Region 3
Dave Montali (Modeling Workgroup Chair), Tetra Tech, West Virginia Department of
Environmental Protection
Mark Bennett (Modeling Workgroup Chair), U.S. Geological Survey (USGS)
Cassandra Davis (Watershed Technical Workgroup Chair), New York State Department of
Environmental Conservation

Dear WQGIT, Modeling Workgroup and Watershed Technical Workgroup Chairs,

Please accept this request and update from the Conowingo WIP (CWIP) Steering Committee (SC) regarding modeling efforts to determine the nutrient reduction efficiency of Conowingo dredging. This request and update is in follow up to the Conowingo modeling approach presented at the <u>July 26, 2021 WQGIT meeting</u>.

Nutrient reductions associated with Conowingo dredging are site-specific and influenced by local hydrodynamic processes, watershed inputs, sediment biogeochemistry, and estuarine fate and transport. It would be inappropriate to derive nutrient reductions from the scientific literature as is typically done by the Chesapeake Bay Program partnership's BMP Expert Panels. As such, and with the concurrence of EPA Chesapeake Bay modelers, the CWIP SC proposed back in July that nutrient reduction efficiencies associated with any Conowingo dredging are best quantified by integrating new or existing Chesapeake Bay and Conowingo modeling tools updated with the most current local data and information regarding Conowingo Pool nutrient bioavailability and geochemistry. We also proposed that oversight and technical evaluation of

the modeling tools was better suited to the charge and technical capacities of the Modeling Workgroup, but that we would also keep the WQGIT closely tied in (also now the Watershed Technical Workgroup) as the body approving loading rate reductions used in the Chesapeake Bay Watershed Model per the <a href="BMP Expert Panel Protocols">BMP Expert Panel Protocols</a>. Please consider the following requests and model status updates with this background and understanding in mind.

Currently, two options are being explored for model development: (1) use of Constellation Energy's (formerly Exelon) Conowingo Pond Mass Balance Model (CPMBM); and, (2) new Corps of Engineers (CoE) modeling through their Planning Assistance to States program. The two options are being pursued in parallel as contingencies. The CWIP SC recently received model documentation for the CPMBM (attached) and are requesting Modeling Workgroup review of this documentation for sufficiency and to provide recommendations on next steps for using the CPMBM to quantify nutrient credits. This is a critical step for deciding whether the CPMBM should continue to be pursued as a viable modeling approach. The CWIP SC is requesting a 60-day timeframe for this review and appropriate coordination with the WQGIT and Watershed Technical Workgroup.

For the CoE Conowingo modeling approach, the CWIP SC worked with EPA modelers to send a draft scope of work to the CoE Baltimore District for comments. The draft scope is being reviewed by the CoE's Engineering Research and Development Center (ERDC). Once we hear back from ERDC we will work with the CoE on next steps and share the draft scope with your teams.

Lastly, and as part of Maryland's Innovative and Beneficial Reuse Pilot, a regression model was also completed to estimate Conowingo nutrient reductions under different dredging scenarios. This is also being submitted for evaluation and comment on how this might fit into the broader modeling approach.

Given this modeling-based approach for Conowingo dredging evaluation which is different from the typical BMP Expert Panel process, we want to ensure that we are proceeding in a proper and transparent way with this effort which would be consistent with the BMP Expert Panel Protocols. We also want to express our strong support for updating the expert panel process and funding the panels. This aligns with the CWIP's goal to advance BMP innovations that will help the partnership meet and sustain our Bay restoration commitments. Thank you in advance for your consideration of these documentation review requests and help in determining appropriate modeling next steps. Please let us know if you have any guestions or concerns.

Best regards,

Jill and Matt

Conowingo WIP Steering Committee

Cc: Michelle Price-Fay, Acting Director, USEPA Chesapeake Bay Program

Lee McDonnel, Chief, Science, Analysis and Implementation Branch, USEPA Chesapeake

Bay Program

ATTACHMENT B: Section 4, draft modeling section from dredging and innovative use pilot study report



# Section 4: Nutrient Flux and Sediment Transport Modeling Results

#### Background & Objectives (4.1)

In 2010, the Chesapeake Bay Total Maximum Daily Load (TMDL) defined numerical reduction targets for sediment, phosphorus, and nitrogen with the goal of meeting these targets by 2025 (EPA, 2010). The TMDL, and the associated modeling that informed the sediment and nutrient reduction targets, assumed that the Conowingo Reservoir would continue to trap sediment and nutrients through 2025. Since that time, and due in large part to ongoing monitoring, considerable advancements have taken place in our understanding of the water quality of the Bay, its connection to various upstream sources of sediment and nutrients, and the delivery of sediment and nutrients throughout the 64,000 square mile watershed. The Chesapeake Bay TMDL included a progress review known as the Mid-Point Assessment that provided an opportunity to review the targeted load reductions and account for new understandings. A preponderance of evidence including bathymetric surveys, mass balance using upstream and downstream monitoring data, and modeling analyses resulted in general agreement among stakeholders that the Conowingo Reservoir is currently in dynamic equilibrium (Zhang et al., 2016). This has been defined as a state associated with equal input and output of materials averaged over long time-periods which has been exhibited in the Conowingo Reservoir since approximately the late 1990s. These findings were integrated into the 2017 Mid-Point Assessment and resulted in the annual addition of 0.26 million pounds per year (0.26-Mlb/yr.) of phosphorus and 6 million pounds of nitrogen (6-Mlb/yr.). This precipitated the development of a Conowingo Watershed Implementation Plan (CWIP), which was finalized July 31, 2021, and outlines the best management practices and strategies to address the increased reductions required to meet the TMDL targets. While the CWIP does not specifically include dredging as a management strategy, it notes that sediment removal needed to be explored further.

Strategic dredging of infill sediments in the Conowingo Reservoir has been considered as a potential sediment and nutrient management strategy since at least 2015 when it was evaluated within the Lower Susquehanna River Watershed Assessment (LSRWA, 2016). Dredging and other Reservoir sediment management strategies are targeted because of the large contribution of sediment and associated nutrients that are delivered to Bay during high-flow events. This has motivated researchers and others to understand the specific contribution that infill sediments play in non-attainment of water quality standards in the upper Bay.

While much of the early research focused on the sediment and nutrient contribution associated with scour events, typically defined as flows that exceed approximately 400,000 cubic feet per second (cfs), recent research has concluded that these large flow events are not necessarily directly responsible for degraded water quality in the Bay. These scour events tend to be relatively infrequent and when they do occur the nutrient and sediment dynamics in the upper Bay tend to trap scoured suspended sediment (Palinkas, 2019 and USACE, 2015). However, if viewed holistically, infill sediments likely influence the timing of both scour-derived and watershed-derived loading to the Bay. As pointed out in Palinkas et al. (2019), a decrease in deposition of watershed sediments within the Reservoir, due to diminished trapping efficiency related to available sediment capacity and particularly with higher energy during flow events, would allow the sediments to remain in suspension and transport downstream to the Bay. The Chesapeake Bay Program's Scientific and Technical Advisory Committee (STAC) identified a similar finding that "net sediment and particulate nutrient deposition behind the Dam are decreasing over a wide range of flows - including flows well below levels typically viewed as scour events" (Linker, 2016). Ongoing research will help determine whether infill- or scour-derived load is contributing to non-attainment of water quality standards in the Bay and at what timescale these impacts take place; however, the conclusion remains that management of infill sediment is an opportunity for intervention.

The objective of this section of the Pilot Project is to develop a planning-level framework to inform decision making for strategic dredging management. The approach includes reviewing and synthesizing available publications and modeling results to identify best practices that can be applied to any strategic dredging management scenario and the development of a planning-level screening tool to approximate the effect of different sediment removal quantities on sediment and nutrient loading and impact towards the needed TMDL reductions. It should be noted that development of novel numerical models or research is beyond the scope of this Pilot Project. There are many well established models and a growing body of research investigating specific aspects of sediment and nutrient dynamics within the Lower Susquehanna River, Conowingo Reservoir, and Chesapeake Bay. Instead, this Pilot Project seeks to understand the implications of this research and current regulatory requirements on potential scenarios for strategic dredge management.

History and Current State of Conowingo Sediment and Nutrient Modeling (4.2)

The establishment of numerical TMDLs and sound planning for implementation of sediment and nutrient reduction measures requires that models are used to predict water quality conditions. As mentioned above, the 2010 TMDL relied upon modeling to develop the required reductions. The model used is referred to as the Chesapeake Bay Environmental Modeling Package (CBEMP). It consists of an Airshed Model, Land Use Change Model, Watershed Model (currently in Phase 7), and Estuary Model. These models are informed by decades of monitoring data resulting in improvements to our understanding of the mechanisms and dynamics associated with sediment and nutrient processes.

In May of 2015, the U.S. Army Corps of Engineers, Baltimore District (USACE) and the Maryland Department of the Environment (MDE) published the Lower Susquehanna River Watershed Assessment (LSRWA). The LSRWA advanced our understanding of the lower Susquehanna River sediment and nutrient dynamics and their consequences. This work included development of two new models: a hydraulic and sediment transport model of the river corridor from Lake Clarke to the Conowingo Reservoir using HEC-RAS and an Adaptive Hydraulics (AdH) model to simulate hydrodynamics and sediment transport of the Conowingo Reservoir to the Susquehanna Flats, the area below the Conowingo Dam. The AdH model scenarios included 1996, 2008, and 2011 Reservoir bathymetries to evaluate different infill conditions and their effect on transport. Both models used the flow period from 2008 to 2011.

The HEC-RAS model is a 1-dimensional model that is effective at simulating scour and deposition within a river corridor over time. This model provided input information for the AdH model including flow and sediment inflows. The AdH model is a more complex 2-dimensional model that can simulate scour and deposition of bed sediment layers where these patterns may not be uniform across the flow path. The AdH model in turn provided inputs to the CBEMP to understand the downstream impacts of different conditions (Figure 1).

Figure 1. Modeling Process



The LSRWA states that the HEC-RAS model uncertainty is primarily associated with its limited capability to simulate transport of cohesive silt and clay soils. The potential outcome of this could be underestimation of both deposition and scour under some circumstances. Stated uncertainties associated with the AdH model include simulating flocculated sediment coming into the Reservoir, simulating scour of larger compacted sediment aggregates, and the ability to simulate dam operations (USACE, 2015).

To address some of the uncertainties associated with the USACE produced HEC-RAS model documented in the LSRWA, in 2016 a new HEC-RAS model was developed by WEST Consultants and funded by Exelon, Inc. This model utilized gage data from 2008 to 2015 and also provided particle size class inputs to the Conowingo Reservoir.

In June 2017, Exelon published the Conowingo Pond Mass Balance Model (CPMBM) documentation. This work sought to answer questions related to the reactivity of scoured and deposited sediment, their chemical changes, and their transport. The CPMBM includes a hydrodynamic and sediment transport model known as ECOMSED and a water quality model developed by HDR, known as RCA. These models improved upon the previous modeling framework by addressing previous model uncertainties associated with diagenesis, hydrodynamics, transport, and dam operations.

The HEC-RAS model of the lower Susquehanna River and the CPMBM were subsequently reviewed and incorporated into the 2017 Mid-Point Assessment and represented enhancements to the CBEMP that were used to inform the Conowingo Watershed Implementation Plan (CWIP).

## Strategic Dredge Water Quality Impact Calculator (4.3)

Strategic dredging is the most direct method to regain the trapping efficiency and associated transport reduction benefits associated with less Reservoir sediment infill. The assumption inherent in the additional reduction requirements identified in the TMDL Mid-Point Assessment is that if infill volume were returned to quantities associated with a state of non-dynamic equilibrium – e.g., with remaining trapping capacity – then the required reductions or a portion thereof would be accomplished. While the reality of achieving water quality standards and the conditions that bring them about are more complicated, and thus have greater associated uncertainty, the underlying relationship between lower

infill volume and improved downstream water quality is built in to the TMDL, planning, and modeling framework.

The relationship between infill and nutrient loading is utilized for evaluating the relative impact of different dredging scenarios on water quality. To understand the infill condition of the Reservoir, the bathymetry or capacity at different points in time must be known or calculated. The Conowingo Reservoir bottom-surface profile was surveyed in 1959/60, 1990, 1993, 1996/7, 2008, 2011, and 2014 (Langland, 2009; and Langland, 2015). This record provides a robust understanding of infill conditions and patterns. This information in combination with depositional rates from long-term monitoring provides the potential for approximating infill conditions for intervening years to be interpolated while understanding that scour/depositional processes are highly variable depending on flow events.

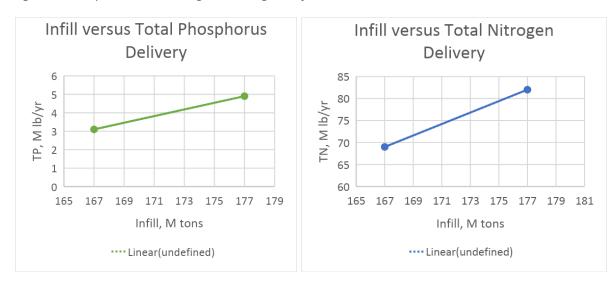
During the Mid-Point Assessment, the Modeling Workgroup (MWG) used the CBEMP to run two scenarios to determine the effect of infill conditions on nutrient loading to the Bay. They used 1995 conditions to represent infill volume that held the previous assumption of remaining capacity and 2010 conditions to represent infill volume at dynamic equilibrium (MWG, 2017). The difference between these two scenarios represents the loading contribution due to infill volume at dynamic equilibrium. It was noted that although scour increases the delivery of particulate nutrients, many of these nutrients are not bioavailable and only reach the Bay during high, infrequent flow events. For this reason, only a portion of these nutrients, those more associated with watershed derived sediments, were assumed to impact water quality and to be relevant to the TMDL.

With the bathymetric and depositional information and the delivered nutrient rates calculated through the MWG efforts, a linear regression can be developed to approximate the effect of different infill volumes on downstream water quality (Table 1). While it is not certain that this relationship is linear, given the nonlinearity of sediment concentrations to flow-events among other uncertainties inherent in a stochastic system, the connection to the CWIP TMDL reductions given the two points of analysis makes this relationship a useful planning level tool to evaluate how infill volume relates to downstream water quality. This relationship is defined in Figure 2 for nutrient delivery rates for the portion of that delivered load that impacts nonattainment of water quality standards and has been incorporated into the TMDL calculations.

Table 1. Infill volume and associated loading for infill states

|      |                    | Delivered Lo    | ad              | TMDL             |                 |                 |
|------|--------------------|-----------------|-----------------|------------------|-----------------|-----------------|
| Year | Infill<br>(M tons) | TP<br>(M Lb/yr) | TN<br>(M Lb/yr) | TSS<br>(M Lb/yr) | TP<br>(M Lb/yr) | TN<br>(M Lb/yr) |
| 1995 | 167                | 3.1             | 69              | 1863             | 0               | 0               |
| 2010 | 177                | 4.9             | 82              | 3217             | 0.26            | 6               |

Figure 2. Phosphorus and Nitrogen Loading vs. Infill Volume



(TP = Total Phosphorous; TN = Total Nitrogen; M = million)

The linear regression above was used to develop a tool that can be used to estimate the impact of infill volume changes such as dredging on yearly nutrient and sediment export and TMDL allocations (Calculator). The Calculator also considers the 1.5 MT/yr of sediment estimated to deposit in the Conowingo Reservoir. Estimates produced by the Calculator are intended as a planning or screening level assessment to understand the potential relative impact of sediment removal. The Calculator does not take into consideration resuspension during dredging or the hydrodynamic impacts of altered bed bathymetry among other characteristics that affect fate and transport of sediment and nutrients within the Reservoir. More sophisticated modeling will be necessary to produce more accurate estimates of the effect of dredging.

The Calculator is designed with tabs that allow dredge quantities to be input in either million cubic yards (MCY) or million tons (MT). This work relied on infill densities as reported and used in previous studies (Langland, 2009) for consistency; however, actual soil densities may vary.

#### Scenario Results and Considerations (4.4)

Table 2 below is from the Calculator and shows results of different dredging quantities and their impact on downstream water quality and specifically their impact on the TMDL allocations outlined in the CWIP.

Table 2. Sample Calculator Results and Annotation

|   |      |              | L            | Delivered Load | TMDL          |              |              |
|---|------|--------------|--------------|----------------|---------------|--------------|--------------|
|   | Year | Infill (MCY) | TP (M Lb/yr) | TN (M Lb/yr)   | TSS (M Lb/yr) | TP (M Lb/yr) | TN (M Lb/yr) |
| ſ | 1995 | 182          | 3.1          | 69             | 1863          | 0            | 0            |
|   | 2010 | 193          | 4.9          | 82             | 3217          | 0.26         | 6            |



|          | Dredge<br>Quantity | Frequency | Loading<br>Reduction | Loading<br>Reduction<br>TN (M | Loading<br>Reduction<br>TSS (M | TMDL<br>Reduction | TMDL<br>Reduction<br>TN (M | As % of<br>6M/yr |
|----------|--------------------|-----------|----------------------|-------------------------------|--------------------------------|-------------------|----------------------------|------------------|
| Scenario | (MCY)              | (yr)      | TP (M Lb/yr)         | Lb/yr)                        | Lb/yr)                         | TP (M Lb/yr)      | Lb/yr)                     | N TMDL           |
| 1        | 1.64               | 1         | 0                    | 0                             | 0                              | 0                 | 0                          | 0%               |
| 2        | 0                  | 1         | 0                    | 0                             | 0                              | 0                 | 0                          | 0%               |
| 3        | 1                  | 1         | 0                    | 0                             | 0                              | 0                 | 0                          | 0%               |
| 4        | 2                  | 1         | 0.05043256           | 0.42643021                    | 50.71865093                    | 0.00814257        | 0.19384442                 | 3%               |
| 5        | 3                  | 1         | 0.21523256           | 1.61633021                    | 174.6186509                    | 0.03194257        | 0.74304442                 | 12%              |
| 6        | 4                  | 1         | 0.38003256           | 2.80623021                    | 298.5186509                    | 0.05574257        | 1.29224442                 | 22%              |
| 7        | 5                  | 1         | 0.54483256           | 3.99613021                    | 422.4186509                    | 0.07954257        | 1.84144442                 | 31%              |
| 8        | 6                  | 1         | 0.70963256           | 5.18603021                    | 546.3186509                    | 0.10334257        | 2.39064442                 | 40%              |
| 9        | 7                  | 1         | 0.87443256           | 6.37593021                    | 670.2186509                    | 0.12714257        | 2.93984442                 | 49%              |
| 10       | 8                  | 1         | 1.03923256           | 7.56583021                    | 794.1186509                    | 0.15094257        | 3.48904442                 | 58%              |

#### Notes:

- -"M" in unit description denotes "million"
- -Strategic dredging is assumed to continue in perpetuity for reductions to be valid. Additional calculations are required if modeling a non-recurring dredge schedule.
- -Frequency specifies the dredge recurrence interval on a yearly basis. For example, if a two is specified, then dredging occurs every 2-years. If 0.5 is specified then dredging occurs twice a year.
- -Infill calculated from Langland, 2009 study.
- -Dredge quantity adjusted to account for 1.5 million ton/yr (1.64 million cubic yeards/yr) depositional inflow (depositional rage from Langland, 2009 and generally cited in many other sources as well)
- -TMDL loads/reductions based on CBP decision in 2017 mid-point assessment modeling
- -Calculator assumes a linear relationship between infill and downstream pollutant loading

Dredging of the Reservoir has been described in the LSRWA and the CWIP as a potential solution to be paired with watershed BMPs that will reduce pollutant inflow to the Conowingo Reservoir. These results support that conclusion by showing that increased dredging of the Reservoir above the rate of depositional inflow is expected to reduce downstream sediment and nutrient loading.

Strategic Dredge Management Best Practices (4.5)

Strategic dredging has many potential benefits that will help reduce sediment nutrient flux to the upper Chesapeake Bay including:

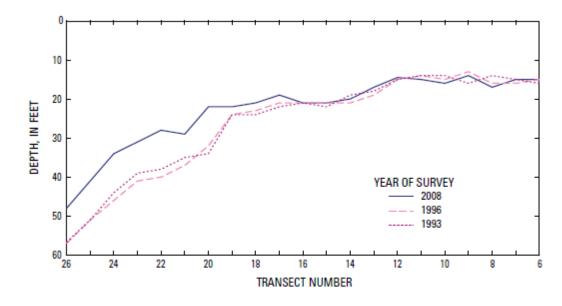
- Physical removal of sediment and attendant nutrients
- Increase in Reservoir storage-capacity
- Decrease in flow velocity
- Increase in particle settling rates
- Increase in sediment deposition
- Decrease of shear stress
- Increase in scour threshold
- Decrease of scour induced transport of sediment and nutrients downstream
- Decrease of suspended load transport of sediment and nutrients downstream
- Increase of sediment trap efficiency

A dredging management strategy should consider the conditions needed to achieve these benefits to derive the greatest reduction in nutrient and sediment flux. Key elements of a dredging management strategy include:

#### Location

Locations within the Reservoir where dredging is to occur and the sequence of multi-year dredging locations could focus areas where deposition has occurred in recent decades (Figures 3 and 4). They could also focus on recent depositional areas (Figure 5). Areas identified as A and B in Figure 4 represent areas of consistent deposition and in the case of Area B, scour during high-flow events. Area A is likely also susceptible to high and moderate flow events due to the narrowing of the Reservoir in this location, which increases velocity and lowers the scour threshold.

Figure 3. Change in depth to bottom surface by transect in Conowingo Reservoir, 1993 to 2008 (Langland,



2009)

Figure 4. Bathymetric survey transects surveyed in Conowingo Reservoir (Langland, 2009)

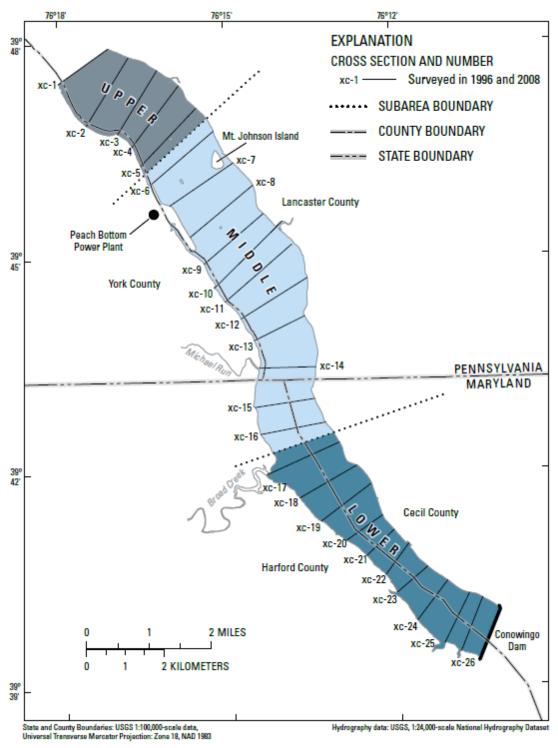
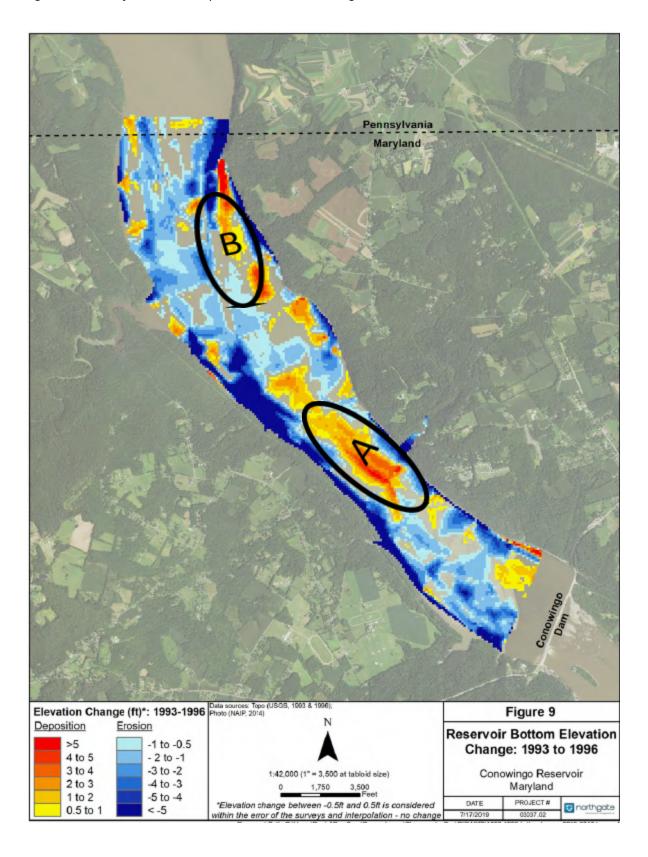


Figure 9. Bathymetric survey transects surveyed in Conowingo Reservoir, lower Susquehanna River.

Figure 5. Areas of consistent deposition in the Conowingo Reservoir.



#### Time of Year

Moderate flow-events, those capable of moving suspended sediment downstream of the dam, occur most frequently during the Spring due to snow melt and moderate to heavy rains associated with the freshet. Higher flow-events occur in the late summer or early fall associated with the Mid-Atlantic hurricane season. Aligning dredge scenarios prior to these events would increase the depositional opportunities resulting from these events and potentially reduce the susceptibility of newly deposited sediments to scour.

#### Volume

To achieve the goal of providing trapping capacity that can capture net volumes of sediment and nutrients that flow downstream, dredge volumes in any year need to be greater than the annual sediment deposition rate of 1.5 MT (1.64 MCY). Results from the regression calculator above indicate that if 3 MCY/yr are dredged in a given year, total Nitrogen delivered to the Bay (TN) will be reduced by 1.616 M lb/yr. Note though that this TN estimate is not the same as the TMDL reduction that would be observed for TN (0.743 M lb/yr). This TMDL reduction would be 12.4% of the EPA's 6 M lb/yr target. Because of the annual deposition rate, doubling sediment removal to 6 MCY/yr would more than double this percentage; the percentage would jump to 40%. Volumes removed annually could be structured in reference to nutrient reductions from a more refined version of the calculator presented here, recognizing practical limitations on maximum potential volume based on a six month window.

#### Depth and Spatial Extent

The profile of Conowingo Reservoir bed sediments is commonly organized by reactivity and referred to as G1, labile; G1, refractory; and G3, inert. The layer of sediment closer to the surface has the greatest bioavailability and thus the greatest impact on downstream water quality. Strategic dredging that focuses on removing a wide rather than deep cross-section of sediment would thereby likely have the

<sup>&</sup>lt;sup>1</sup> This is because "delivered load" reflects nutrients that reach the Bay, whereas TMDL reductions are calculated based on amounts of nutrients considered to be bioavailable after passing through tidal regimes, SAV beds, and biogeochemical conversions that render some nutrients unavailable.

greatest impact by removing the most reactive (labile) layer. Further modeling should take these parameters into account.

#### Time Interval

Given the time of year of the two higher flow conditions, dredging prior to the Spring will reduce the impact of transport during moderate storms in the Spring and potentially reduce the downstream impact of scour events in the late summer.

#### Conclusions and Implications for Further Investigation

This work provides planning or screening level estimates of the potential impact of a dredging management strategy. For example, screening level results include that dredging 3 MCY/yr is expected to provide a TMDL reduction for TN of 12% (vs. the EPA-stated annual need to reduce 6 M lb/yr), and dredging 6 MCY/yr is expected to raise this contribution to 40% TMDL reduction. However, it is expected that more precise measurements of pollutant reduction would be needed prior to implementation of a full dredging approach. Currently, and as described in Section 4.2, the Chesapeake Bay Program has used the CPMBM developed by HDR and funded by Exelon, Inc. to produce input into the CBEMP. It is recommended that a modified modeling approach be used to generate more precise pollutant reduction quantities. It is also expected that more focused modeling may show a greater TMDL reduction than calculated in this study, because dredge activities would be tailored to 1) locations where deposition is most likely to occur and 2) times of year when resuspension would be minimized. To model changes that dredging would cause, Conowingo Reservoir bathymetry would need to be modified in the dredging locations and updated at the frequency outlined in the dredging scenario. G-fraction reactivity differences should also be incorporated as described above. Such a modeling effort should be conducted in a manner consistent with CPMBM and able to be incorporated into forecasting models using CBEMP.

Also, because the CPMBM model has no feedback between changes in bathymetry resulting from sediment deposition/erosion and Reservoir hydrodynamics, it would be unable to represent the process by which dredged areas of different dimensions fill with sediment over multi-year simulations, during which the depth of the dredged area is decreased and sediment trapping efficiency is gradually reduced. Further, the HDR model has no windwave resuspension, which is important because sediment seasonally deposited at the upper end of the Reservoir is redistributed into deeper parts of the Reservoir during

high flow events. Future comparisons of Reservoir dynamics under different dredging scenarios should take this resuspension into account. A strategic dredging modeling effort that addresses these issues could produce more refined results than this project; results may include slightly different TMDL implications from those found here.