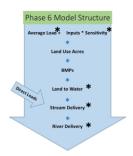
Testing watershed properties as candidate predictors of long-term average streamflow

Isabella Bertani, Gopal Bhatt, Gary Shenk

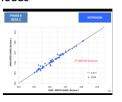
Modeling Workgroup Quarterly Review 7/6/2021

CalCAST Hydrology Model Development

Average annual streamflow (Q) is the difference of Rainfall and Actual Evapotranspiration (AET), where AET can be estimated from Potential Evapotranspiration (PET) and/or other watershed properties.



Calibration of metaparameters to spatial loads



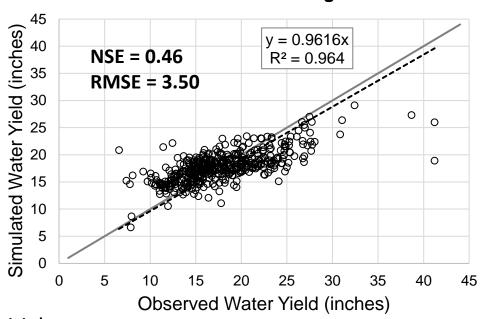
$$Q = \sum_{\substack{upstream \\ geography}} Precipitation - PET \times f_{LU} \times Fn(watershed properties)$$

e.g.,
$$Fn(watershed\ properties) = a + f_w \times Wetness + f_s \times Slope$$

Results (CalCAST Hydrology – an initial, simplified prototype)

After Calibration

Water Yield for 453 Monitoring Stations



$$NSE = 1 - \frac{\sum_{i=1}^{n} (pred_i - obs_i)^2}{\sum_{i=1}^{n} (obs_i - \overline{obs})^2}$$

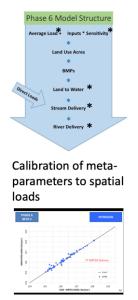
$$RMSE = \sqrt{\sum_{i=1}^{n} \frac{\left(obs_{i} - pred_{i}\right)^{2}}{n}}$$

Tests with different initial parameters, objective functions, etc. were conducted.

Stats summary

84% of all station within ±25% 89% of ≥ 50cfs P6 stations 77% of new stations **Goal**: Test candidate watershed properties for potential inclusion in CalCAST to improve average annual streamflow prediction

Broader goal: transition from P6 hydrology calibration largely based on adjusting PET by county to calibration based on mechanistically plausible / management-relevant properties (land use, watershed characteristics, climate...)



$$Q = \sum_{\substack{upstream \\ geography}} Precipitation - PET \times f_{LU} \times Fn(watershed properties)$$

e.g., $Fn(watershed\ properties) = a + f_w \times Wetness + f_s \times Slope$

Climate (long-term average)

Air Temperature Percent of precipitation as snow

Geology

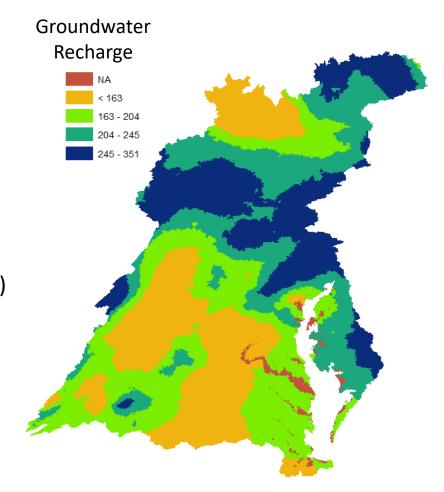
Bedrock permeability classes
Hunt geology
Principal aquifers and rock types (USGS, 2003)
Olson geology types (Olson, 2014)
Soller surficial materials (Soller et al., 2009)

Generalized lithologic classes (Anning & Ator, 2017)

Dominant geology type (Reed & Bush, 2001)

Topography

Catchment mean slope Catchment mean elevation Stream mean slope Stream length/Catchment Area



Land use

% of 12 land use classes (crop, forest, impervious, etc..)

Hydrology

Base Flow Index

Contact time

RUSLE R factor

Groundwater recharge

Saturation overland flow as % of streamflow

Topographic Wetness Index

Depth to water table

Infiltration-excess overland flow as % of streamflow

Enhanced vegetation index

Soil

Texture (% clay, sand, silt)

Hydrologic groups

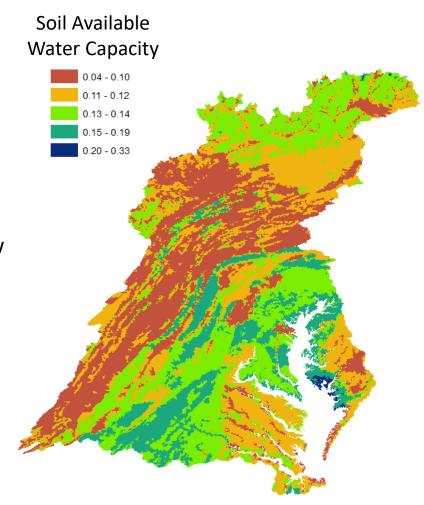
Permeability

Moisture storage

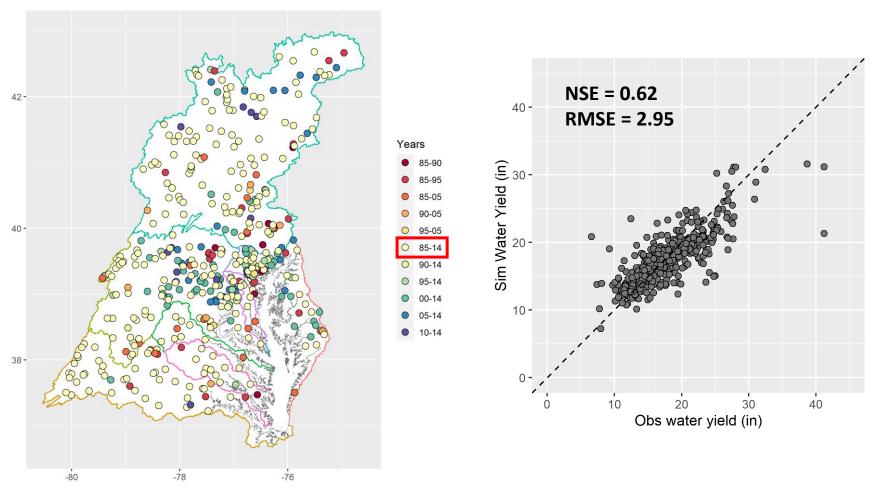
Available water capacity

Percent organic material

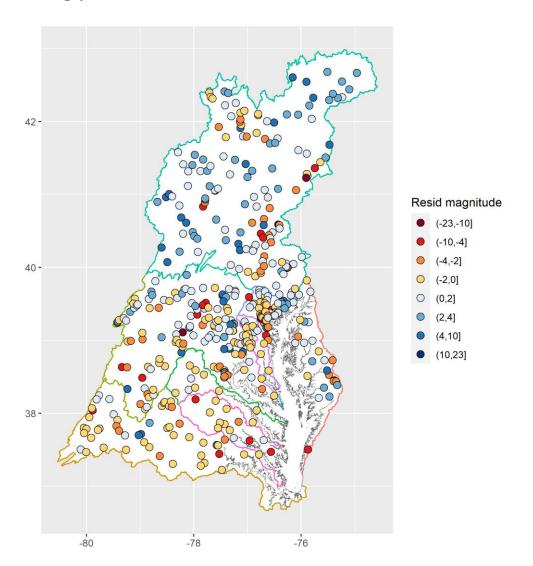
Rock depth



First step: Stations have different periods of record. We matched PCP and PET inputs to years of record at each station (we previously used 1991-2000 at all stations just to get the code working)

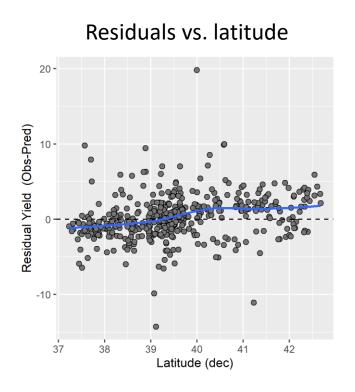


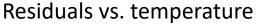
Second step: analysis of residuals (OBS – SIM) to look for patterns that may point to missing predictors

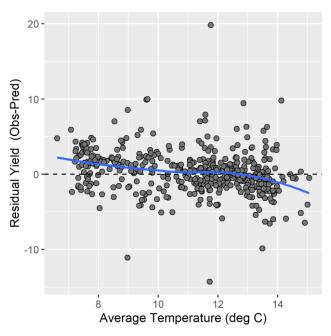


Tendency to
underpredict (blue
shades) streamflow in
upper portion of
watershed and
overpredict in lower
portion (red shades)

Second step: analysis of residuals (OBS – SIM) to look for patterns that may point to missing predictors

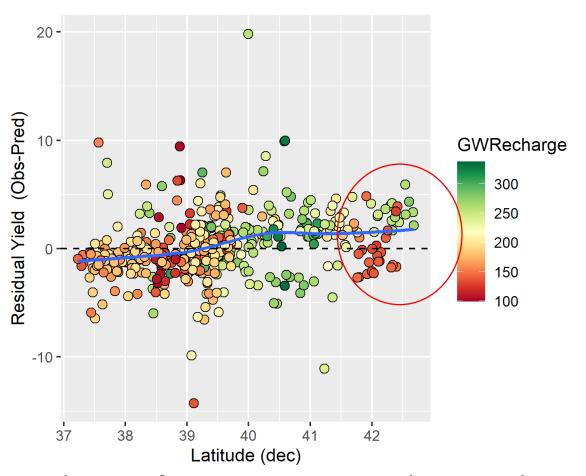






Inclusion of a variable like latitude or temperature may help remove this north-south spatial pattern in residuals

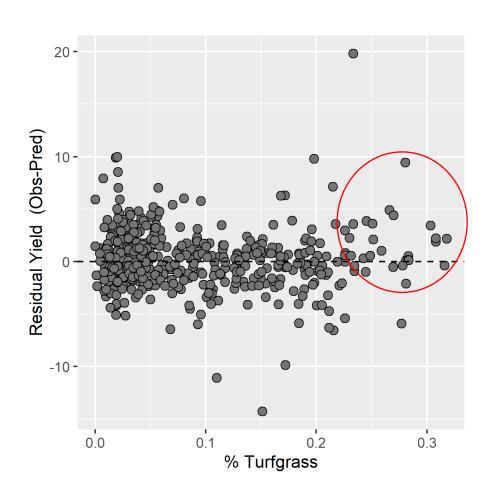
Second step: analysis of residuals (OBS – SIM) to look for patterns that may point to missing predictors



In upper portion of watershed, some tendency to overpredict in low gw recharge area and underpredict in high gw discharge area

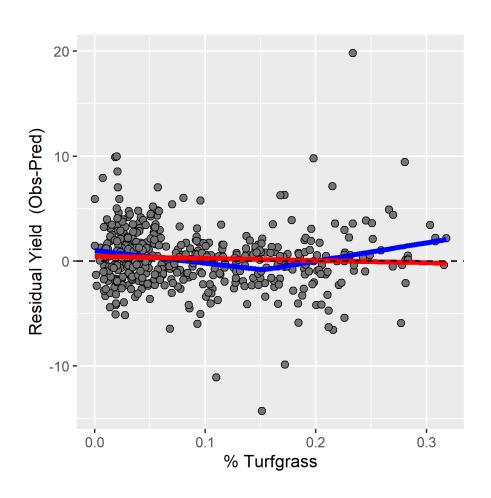
Inclusion of an interaction term between latitude and gw recharge may help capture this pattern

Second step: analysis of residuals (OBS – SIM) to look for patterns that may point to missing predictors



Some tendency to underpredict at very high levels of % Turfgrass (high % of turfgrass tend to co-occur with high % of impervious surfaces)

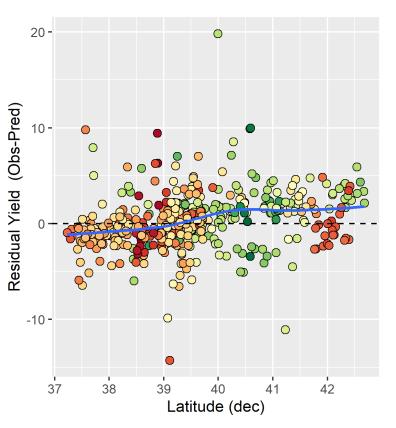
Second step: analysis of residuals (OBS – SIM) to look for patterns that may point to missing predictors



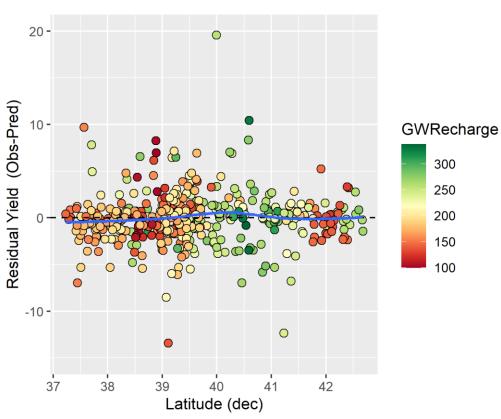
A piecewise regression
(blue line) better
captures this apparent
non-linear pattern than a
simple linear regression
(red line).

Inclusion of %Turfgrass as a piecewise regression term in the model may help capture this pattern

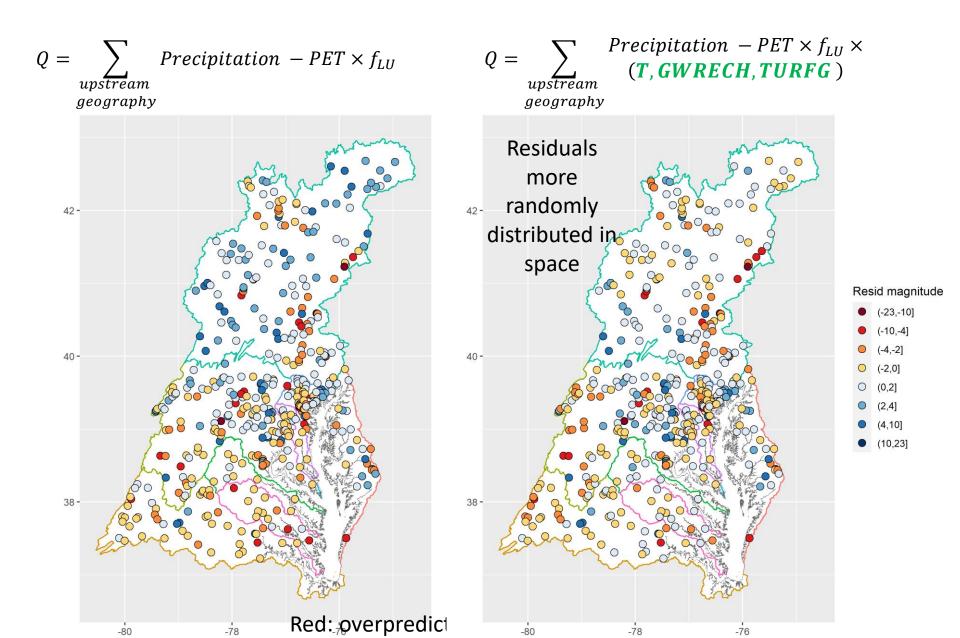
$$Q = \sum_{\substack{upstream \\ geography}} Precipitation - PET \times f_{LU}$$



$$Q = \sum_{\substack{upstream \\ geography}} \frac{Precipitation - PET \times f_{LU} \times}{(T, GWRECH, TURFG)}$$

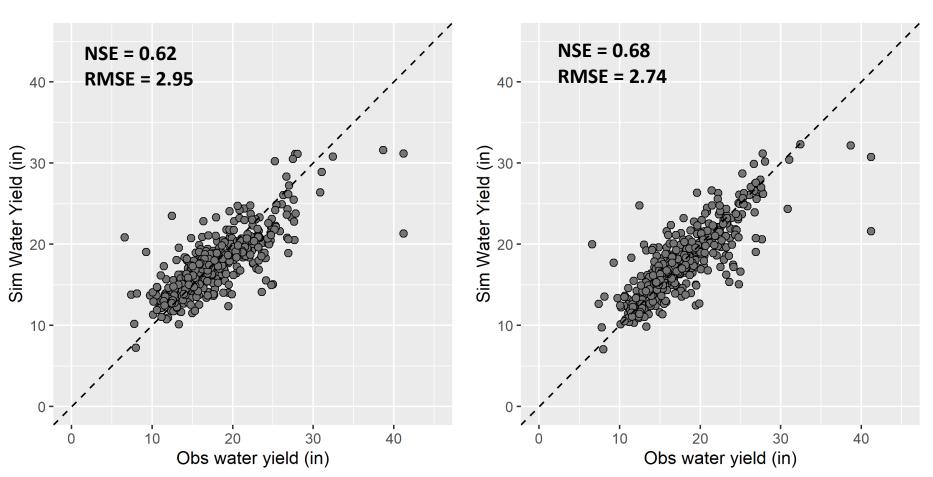


Substantially decreased overprediction at low latitudes and underprediction at high latitudes/high gw recharge.

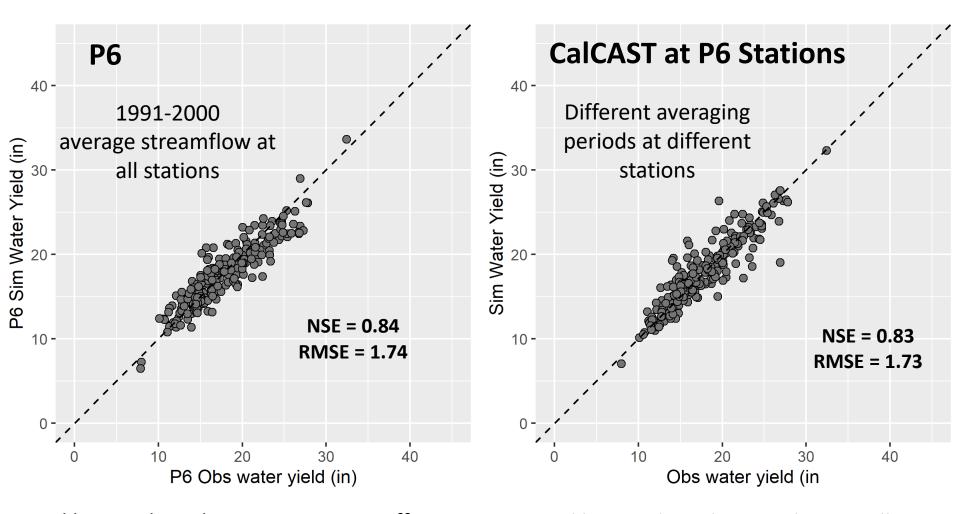


$$Q = \sum_{\substack{upstream \\ geography}} Precipitation - PET \times f_{LU}$$

$$Q = \sum_{\substack{upstream \\ geography}} \frac{Precipitation - PET \times f_{LU} \times}{(T, GWRECH, TURFG)}$$



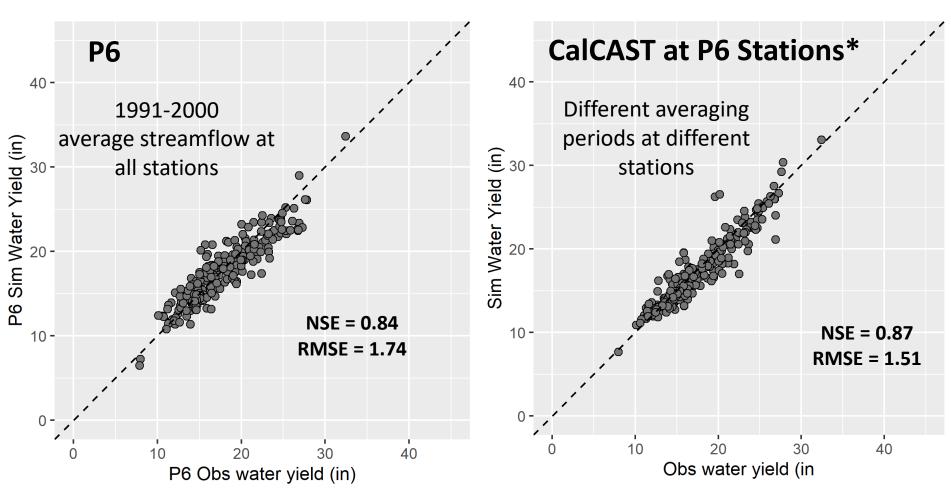
Comparison with P6 hydrology



Calibration based on varying PET coefficient by county

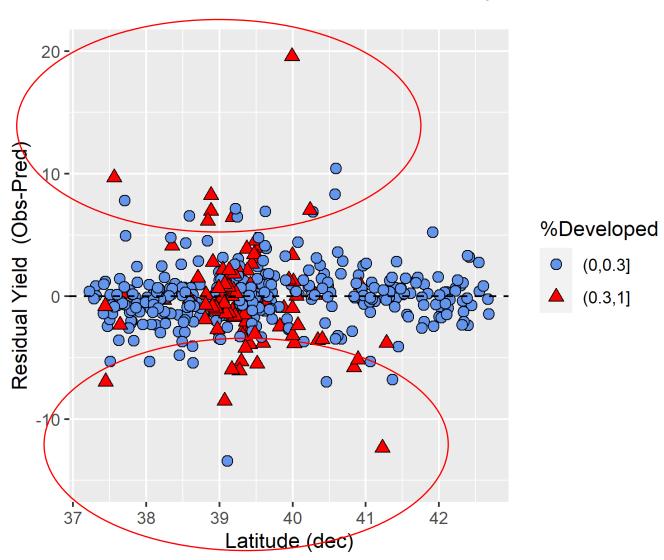
Calibration based on mechanistically meaningful watershed properties

Comparison with P6 hydrology



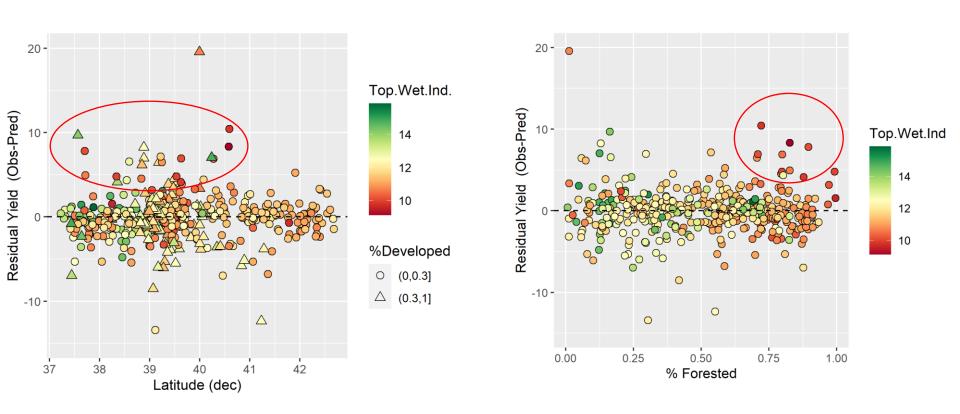
*Replacing **PCP** and **PET** inputs with **RUNOFF** predicted by relatively simple water balance model (Wolock & McCabe, 2018)

Room for improvement



Several of the stations where the model performs the worst have relatively high % developed land (triangles)

Room for improvement



Several of the stations where the model performs the worst (of those with low % developed) exhibit some of the largest % forested land and some of the lowest Topographic Wetness Index values

Room for improvement

$$Q = \sum_{\substack{upstream\\geography}} \frac{Precipitation - PET \times f_{LU} \times (T, GWRECH, TURFG)}{(T, GWRECH, TURFG)}$$

 \bigcirc

0.50

% Forested

0.75

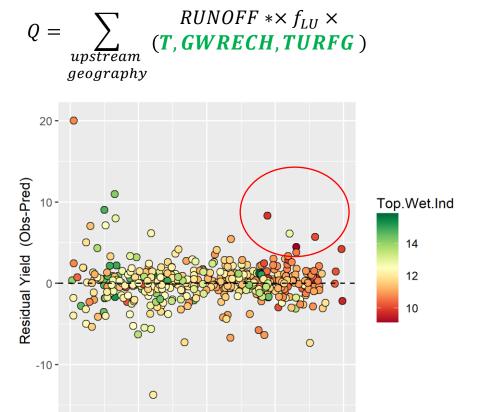
1.00

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0.25

-10 **-**

0.00



0.25

0.50

% Forested

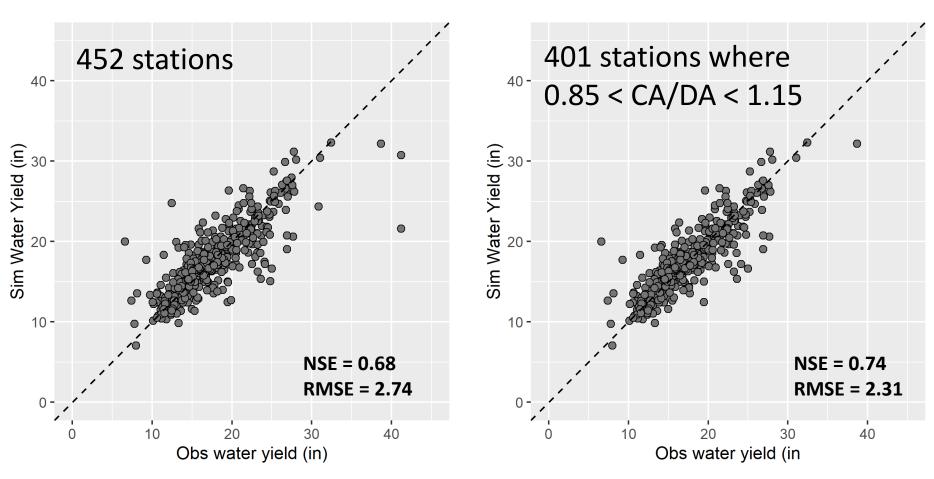
0.00

*Replacing **PCP** and **PET** inputs with **RUNOFF** predicted by relatively simple water balance model (Wolock & McCabe, 2018)

0.75

1.00

Room for improvement



CA: station's catchment area estimated as sum of all upstream NHDPlus catchments DA: station's upstream drainage area provided by USGS

Conclusions

We will continue to explore candidate predictors that may help base hydrology calibration on mechanistically plausible / management-relevant properties of the watershed

Most likely a hybrid calibration approach will be needed (combination of relevant watershed properties and P6-like approach)

Comparison of Modeled and Monitored Nutrient Trends

Isabella Bertani, Gopal Bhatt, Gary Shenk, and the Factors Team

Modeling Workgroup Quarterly Review 7/6/2021

Comparing modeled and monitored nutrient trends

Management Question

- TMDL: Implement the practices by 2025 that will eventually lead to meeting water quality standards
- CAST prediction: What is the long-term load resulting from a given state of the watershed (land use, point sources, management actions, etc)
- WRTDS- flow normalized loads: based on a moving relationship between flow and concentration, how do loads change over time if annual flow is the same
- Science question: How do we use monitoring data to validate the predictions of CAST

Comparing modeled and monitored nutrient trends

Two major objectives:

- Help understand and communicate where and why monitoring data and CAST do not match and how those differences can be reconciled
- Inform future refinements of the watershed model

Comparing modeled and monitored nutrient trends

CAST vs WRTDS_FN is not an "apples-to-apples" comparison

CAST and WRTDS Differences

- Unrealistic expectations
 - Implementation amount
 - BMP effects
- Lag times
 - Implementation / maturation of BMPs
 - Groundwater
 - Soil equilibration
- Insufficient Monitoring
 - Quantified as uncertainty in WRTDS trends
- Competing effects
 - Conowingo
 - · Climate change
 - Weather cycle effects



Factors driving nutrient trends in streams of the Chesapeake Bay watershed

Scott W. Ator¹ [□] | Joel D. Blomquist¹ [□] | James S. Webber² [□] | Jeffrey G. Chanat² [□]

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³ USGS, 1730 East Parham Rd., Richmond, VA 22228 USA

Correspondence Scott W. Ator, USGS, 5522 Research Park Dr., Baltimore, MD 21228, USA. Email: swator@usgs.gov

Assigned to Associate Editor Yongshan

Abstract

Despite decades of effort toward reducing nitrogen and phosphorus flux to Chesapeake Bay, water-quality and ecological responses in surface waters have been mixed. Recent research, however, provides useful insight into multiple factors complicating the understanding of nutrient trends in bay tributaries, which we review in this paper, as we approach a 2025 total maximum daily load (TMDL) management deadline. Improvements in water quality in many streams are attributable to management actions that reduced point sources and atmospheric nitrogen deposition and to changes in climate. Nutrient reductions expected from management actions, however, have not been fully realized in watershed streams. Nitrogen from urban nonpoint sources has declined, although waterquality responses to urbanization in individual streams vary depending on predevelopment land use. Evolving agriculture, the largest watershed source of nutrients, has likely contributed to local nutrient trends but has not affected substantial changes in flux to the bay. Changing average nitrogen yields from farmland underlain by carbonate rocks, however, may suggest future trends in other areas under similar management, climatic, or other influences, although drivers of these changes remain unclear. Regardless of upstream trends, phosphorus flux to the bay from its largest tributary has increased due to sediment infill in the Conowingo Reservoir. In general, recent research emphasizes the utility of input reductions over attempts to manage nutrient fate and transport at limiting nutrients in surface waters. Ongoing research opportunities include evaluating effects of climate change and conservation practices over time and space and developing tools to disentangle and evaluate multiple influences on regional water quality.

1 | INTRODUCTION

Recent efforts toward reducing nutrient flux to Chesapeake Bay from its watershed have been insufficient to meet water-quality and ecological standards in the bay (Chesapeake Bay Program, 2018a; Kleinman et al., 2019; Linker,

Abbreviations: SPARROW, SPAtially Referenced Regressions On Watershed attributes; TMDL, total maximum daily load.

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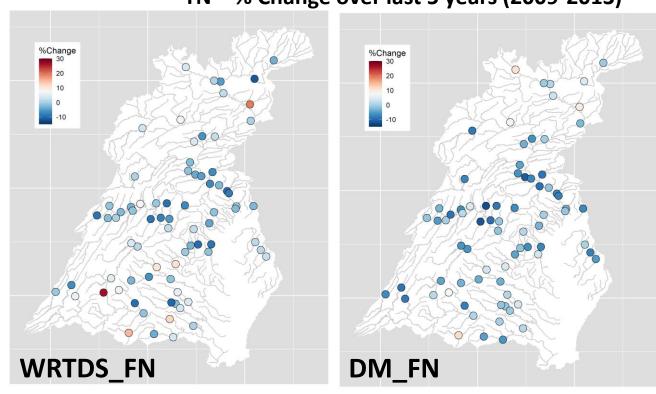
Environ Qual. 2020;49:812-834.

Dynamic watershed model (DM): same inputs as CAST but accounts for (gw) lag times, varying hydrology (and flow-normalization effect on it), non-stationary watershed response to changing conditions

 Explore patterns in differences between WRTDS_FN and DM_FN trends as a function of "unrealistic expectations"

TN – % Change over last 5 years (2009-2013)

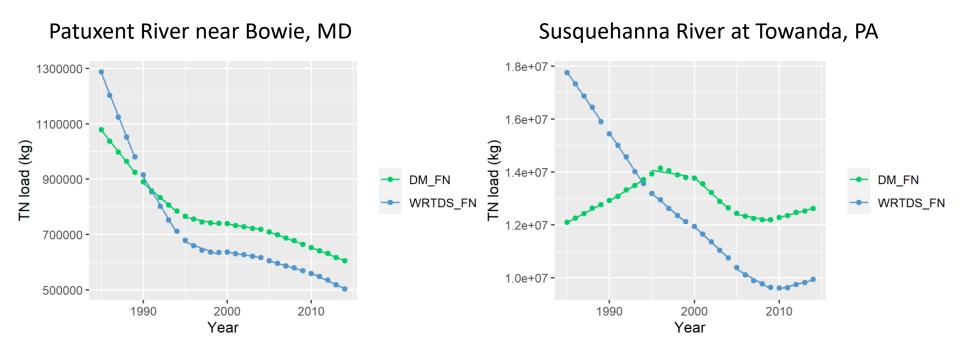
- Land Use
- Nutrient inputs
- BMP type/level
- Watershed characteristics
- Time period
- ...



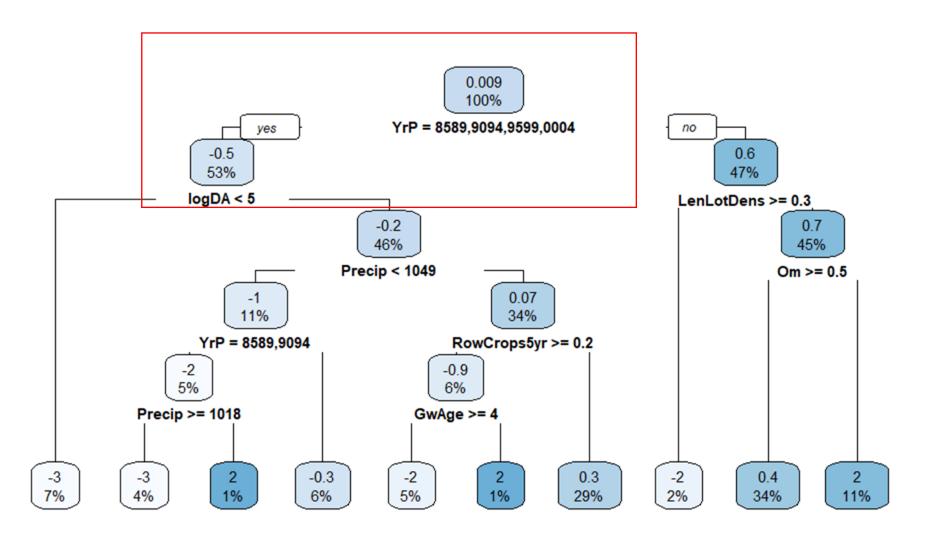
Example of exploratory analysis

Q: where and when are trends in WRTDS_FN and DM_FN most similar/different? Why is that?

Example response variable: Slope of WRTDS_FN vs. DM_FN for consecutive 5-year periods (85-89, 90-94, 95-99, 00-04, 05-09,10-14)



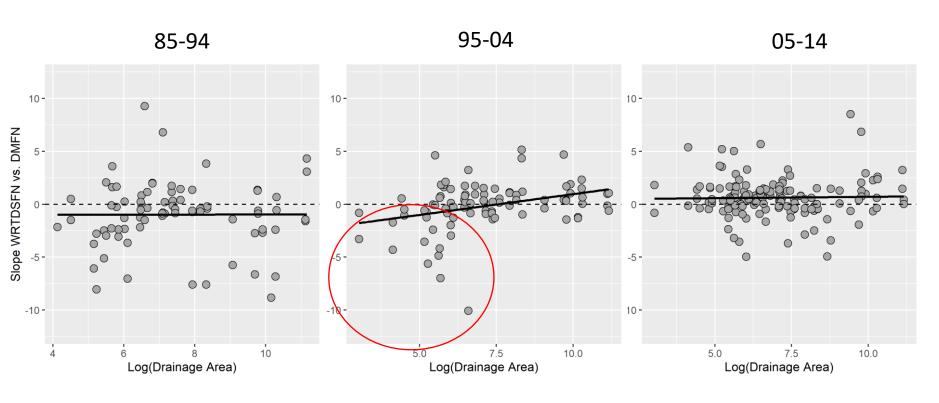
Example of exploratory analysis



^{*}**YrP** = 5-year time period (85-89, 90-94, 95-99, 00-04, 05-09,10-14)

^{*}logDA = log(Station Drainage Area)

Slope of WRTDS_FN vs. DM_FN and Drainage Area

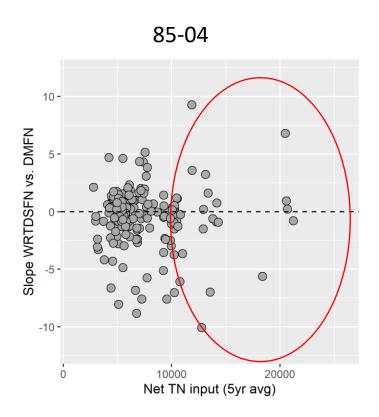


Larger scatter compared to other periods – strongest disagreement between WRTDS and DM

Stations with smaller drainage areas show strongest disagreement

The same is not true in later years

Slope of WRTDS_FN vs. DM_FN and TN inputs

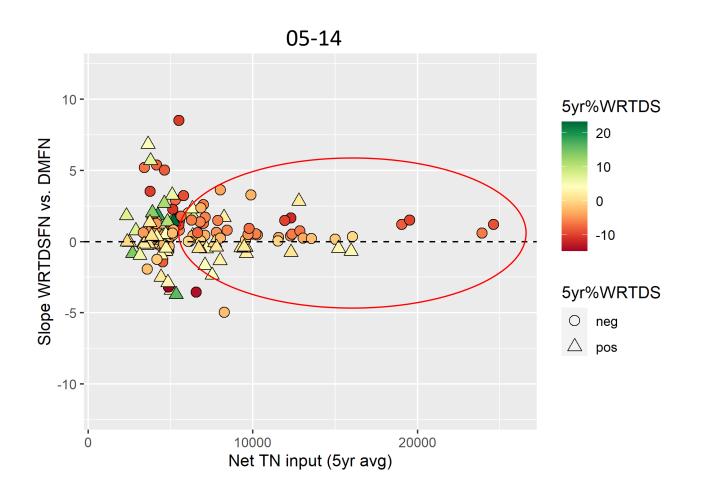


05-14 10 -Slope WRTDSFN vs. DMFN -10 **-**10000 20000 Net TN input (5yr avg)

Stations with larger inputs do not show better agreement in earlier years

Better agreement (slopes closer to 1) at stations with larger inputs in later years

Slope of WRTDS_FN vs. DM_FN and TN inputs



At stations with larger inputs, stations where DM and WRTDS show opposite trends (slope < 0) tend to be those that show an increase in WRTDSFN (triangles) in later years

Summary

DM_FN and WRTDS_FN exhibit the largest disagreement at stations with smaller drainage area in the middle of the period of record, but the same is not true in earlier and later years

In later years, DM_FN and WRTDS_FN exhibit good agreement at stations with high TN inputs and where WRTDS_FN exhibits a negative trend. At stations with high TN loads where WRTDS_FN exhibits a positive trend in later years, DM_FN has opposite trend

In later years, most disagreement at stations with low TN inputs, working on understanding why